

Water Research

Analytical investigation of cannabis biomarkers in raw urban wastewater to refine consumption estimates

--Manuscript Draft--

Manuscript Number:	
Article Type:	Research Paper
Keywords:	Δ^9 -tetrahydrocannabinol; THC metabolites; liquid-liquid extraction; LC-MS/MS; suspended solids; Wastewater-based epidemiology
Corresponding Author:	Lubertus Bijlsma University Jaume I Castellon, SPAIN
First Author:	Marina Celia Campos-Mañas
Order of Authors:	Marina Celia Campos-Mañas Natan Van Wichelen Adrian Covaci Alexander L.N. van Nuijs Christoph Ort Frederic Béen Sara Castiglioni Félix Hernández Lubertus Bijlsma
Abstract:	<p>Wastewater analysis of Δ^9-tetrahydrocannabinol (THC) biomarkers can provide essential information on trends in cannabis consumption. Although analysis is mostly focussed on the aqueous phase, previous studies have illustrated the need of improving the measurements of raw influent wastewater (IWW) considering also suspended solids. Especially for cannabis biomarkers, a substantial part of them is expected to be found to solids in wastewater due to their hydrophilic character. However, it remains open to which extent trend estimates might be affected by solely analysing the liquid phase. To investigate this aspect, robust analytical methodologies are required to measure both the liquid and solid phase of IWW. In this work we firstly tested liquid-liquid extraction (LLE) for THC and its major metabolites (THC-OH, and THC-COOH). Using LLE, no filtration or centrifugation step was required and the three analytes were extracted from both the liquid and the solid phase simultaneously. In parallel, the same samples were centrifuged and analysed as follows: the liquid phase separately by both LLE and solid phase extraction (SPE) and the suspended solids by solid-liquid extraction (SLE). The separate analysis of the liquid and solid phase in different IWW samples revealed that a significant amount of cannabis biomarkers (ranged from 42 to 90%) could be found in the suspended solids. In addition, the total amount of cannabis biomarkers observed by analysing raw IWW on the one hand, and by separate analysis of the liquid and the solid phase on the other hand, was in good agreement. Data from this study show that the sole analysis of the liquid phase would lead to a notable underestimation (50%) of cannabis biomarkers determination in IWW.</p>
Suggested Reviewers:	Cobus Gerber University of South Australia cobus.gerber@unisa.edu.au Daniel Burgard University of Puget Sound dburgard@pugetsound.edu Imma Ferrer University of Colorado imma.ferrer@colorado.edu

	Rosario Rodil University of Santiago de Compostela: Universidade de Santiago de Compostela rosario.rodil@usc.es
--	---

	Richard Bade University of Queensland Faculty of Biological and Chemical Sciences: The University of Queensland Faculty of Science r.bade@uq.edu.au
--	--

Water Research
Editorial Office

Castellon, 4th of March 2022

Dear Editor,

Please, find enclosed our paper entitled “Analytical investigation of cannabis biomarkers in raw urban wastewater to refine consumption estimates”, which we submit for your consideration for publication in Water Research

Wastewater analysis of Δ 9-tetrahydrocannabinol (THC) biomarkers can provide essential information on trends in cannabis consumption in wastewater-based epidemiology (WBE) studies. However, it remains unclear to which extent estimates may be affected by solely analysing the liquid phase.

In this study, we analyzed Δ 9-tetrahydrocannabinol (THC) and its major metabolites (THC-OH and THC-COOH) from both the liquid and the solid phase of wastewater samples. This information was complemented by results obtained from the analyte stability study in samples. Data from this paper revealed that a significant amount of cannabis biomarkers (ranged from 42 to 90%) were found in suspended solids, and therefore the sole analysis of the liquid phase would lead to a notable underestimation (50%) of cannabis biomarkers present in influent wastewater. Thus, an important knowledge gap in the application of WBE for cannabis use estimation has been identified and carefully studied in this work.

The results obtained in this work allow a better understanding of the occurrence and partition of cannabis biomarkers between the liquid phase and the suspended solids in influent wastewater, and will help to improve surveillance of cannabis consumption at future events. This study is an important piece of the puzzle that give more insight in how to use WBE as a tool to monitor cannabis consumption. In addition, several key issues have been identified for future WBE research, i.e. assessing sampling uncertainty, evaluation of the cannabis biomarkers behavior during in-sewer transport, obtaining accurate urinary and fecal excretion rates.

We feel our paper fits well for publication in Water Research, and will be of interest especially for those researchers working on analysis of drugs in (wastewater).

We look forward to your evaluation of our manuscript. Please notify me if I may be of assistance.



Yours sincerely,

Dr. Lubertus Bijlsma
Analytical Chemistry and Public Health
Research Institute for Pesticides and Water, University Jaume I, Castellón, SPAIN.
bijlsma@uji.es
Tel: +34 964 387452; Fax: +34 964 387368

Details

Manuscript title of the research article presented:

Analytical investigation of cannabis biomarkers in raw urban wastewater to refine consumption estimates

Words: 4369 (for the introduction, methods, results and discussion)

Figures: 4

Tables: 4

Supporting information:

Figures: 4

Tables: 5

Authors and affiliates:

Marina Celia Campos-Mañas^a, Natan Van Wichelen^b, Adrian Covaci^b, Alexander L.N. van Nuijs^b, Christoph Ort^c, Frederic Béen^d, Sara Castiglioni^e, Félix Hernández^a, Lubertus Bijlsma^{a*}

^a Environmental and Public Health Analytical Chemistry, Research Institute for Pesticides and Water (IUPA), University Jaume I, Castellón, Spain

^b Toxicological Centre, University of Antwerp, Antwerp, Belgium

^c Swiss Federal Institute of Aquatic Science and Technology (Eawag), Dübendorf, Switzerland

^d KWR Water Research Institute, Nieuwegein, the Netherlands

^e Istituto di Ricerche Farmacologiche Mario Negri IRCCS, Department of Environmental Health Sciences, Milan, Italy

Corresponding author:

LUBERTUS BIJLSMA

bijlsma@uji.es

Tel: +34 964 387452

Fax: +34 964 387368

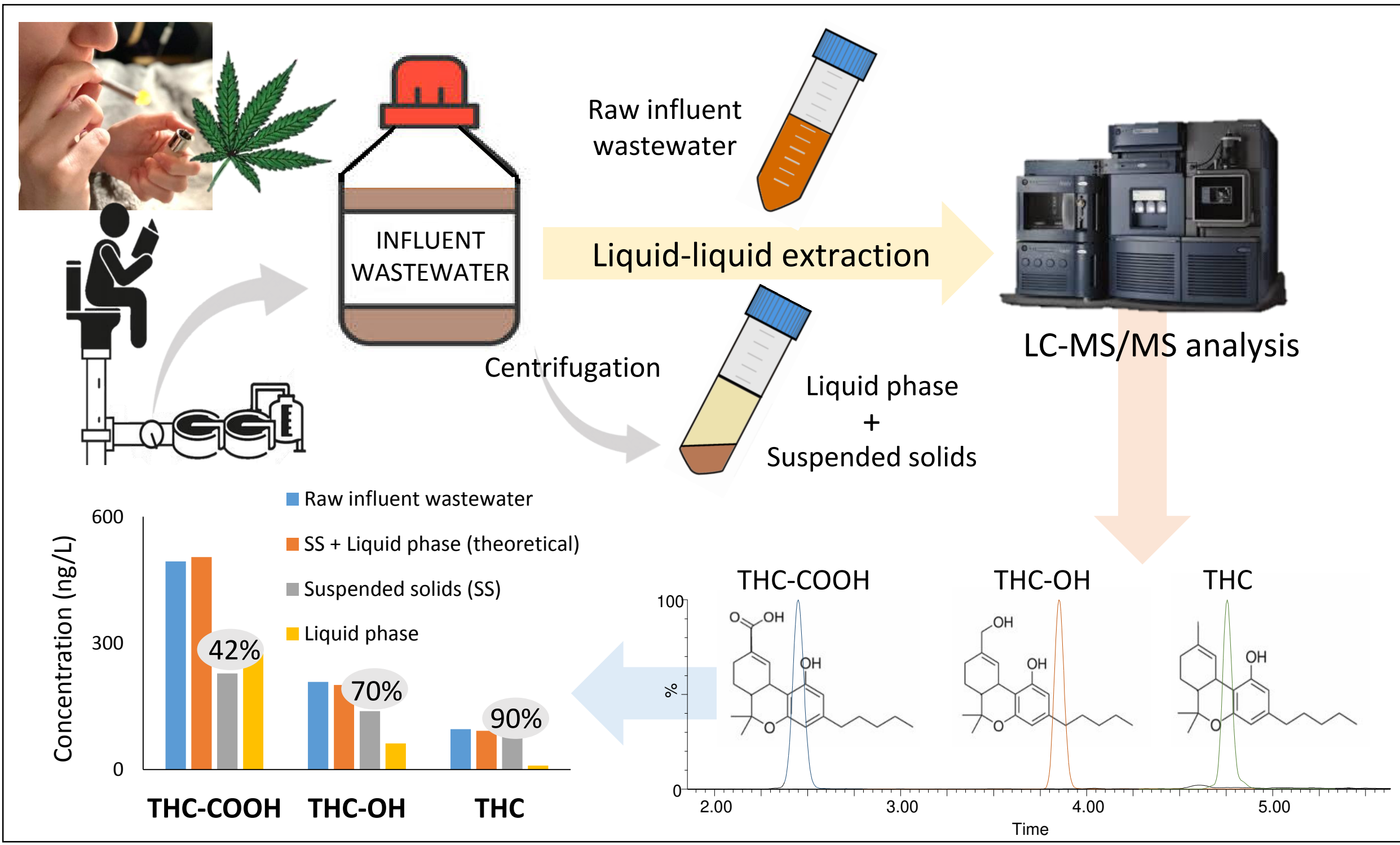
Declaration of interests

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

The authors declare the following financial interests/personal relationships which may be considered as potential competing interests:

HIGHLIGHTS

- Improvements in measuring cannabis biomarkers in raw influent wastewater
- Liquid-liquid extraction as an alternative to conventional solid phase extraction
- Significant amount of cannabis biomarkers is in the suspended solids of wastewater
- Partition of cannabis biomarker in the liquid and solid phase of influent wastewater
- More knowledge on how to use wastewater-based epidemiology for monitoring cannabis



Analytical investigation of cannabis biomarkers in raw urban wastewater to refine consumption estimates

Marina Celia Campos-Mañas^a, Natan Van Wichelen^b, Adrian Covaci^b, Alexander L.N.
van Nuijs^b, Christoph Ort^c, Frederic Béen^d, Sara Castiglioni^e, Félix Hernández^a,
Lubertus Bijlsma^{a*}

*^aEnvironmental and Public Health Analytical Chemistry, Research Institute for
Pesticides and Water (IUPA), University Jaume I, Castellón, Spain*

^bToxicological Centre, University of Antwerp, Antwerp, Belgium

*^cSwiss Federal Institute of Aquatic Science and Technology (Eawag), Dübendorf,
Switzerland*

^dKWR Water Research Institute, Nieuwegein, the Netherlands

*^eIstituto di Ricerche Farmacologiche Mario Negri IRCCS, Department of
Environmental Health Sciences, Milan, Italy*

*Corresponding author

E-mail address: bijlsma@uji.es, Tel: +34 964 38 7382

1 **Abstract**

2 Wastewater analysis of Δ^9 -tetrahydrocannabinol (THC) biomarkers can provide essential
3 information on trends in cannabis consumption. Although analysis is mostly focussed on
4 the aqueous phase, previous studies have illustrated the need of improving the
5 measurements of raw influent wastewater (IWW) considering also suspended solids.
6 Especially for cannabis biomarkers, a substantial part of them is expected to be found to
7 solids in wastewater due to their hydrophilic character. However, it remains open to which
8 extent trend estimates might be affected by solely analysing the liquid phase. To
9 investigate this aspect, robust analytical methodologies are required to measure both the
10 liquid and solid phase of IWW. In this work we firstly tested liquid-liquid extraction
11 (LLE) for THC and its major metabolites (THC-OH, and THC-COOH). Using LLE, no
12 filtration or centrifugation step was required and the three analytes were extracted from
13 both the liquid and the solid phase simultaneously. In parallel, the same samples were
14 centrifuged and analysed as follows: the liquid phase separately by both LLE and solid
15 phase extraction (SPE) and the suspended solids by solid-liquid extraction (SLE). The
16 separate analysis of the liquid and solid phase in different IWW samples revealed that a
17 significant amount of cannabis biomarkers (ranged from 42 to 90%) could be found in
18 the suspended solids. In addition, the total amount of cannabis biomarkers observed by
19 analysing raw IWW on the one hand, and by separate analysis of the liquid and the solid
20 phase on the other hand, was in good agreement. Data from this study show that the sole
21 analysis of the liquid phase would lead to a notable underestimation (50%) of cannabis
22 biomarkers determination in IWW.

23

24 **Keywords:** Δ^9 -tetrahydrocannabinol, THC metabolites, liquid-liquid extraction, LC-
25 MS/MS, suspended solids, wastewater-based epidemiology

26 **1. Introduction.**

27 Cannabis is worldwide the most commonly consumed illicit drug (European Monitoring
28 Centre for Drugs and Drug Addiction (EMCDDA), 2017; “United Nations Office on
29 Drugs and Crime,” n.d.). The European drug report 2021 (European Monitoring Centre
30 for Drugs and Drug Addiction (EMCDDA), 2021) indicates that cannabis is an
31 established drug, and new forms of cannabis with high Δ^9 -tetrahydrocannabinol (THC)
32 content are now available on the illicit market such as cannabis oil/liquid taken orally or
33 for vaping, edibles, drinks, concentrates (e.g., wax, shatter, budder) or tinctures (e.g.,
34 concentrated amounts ingested orally or taken under the tongue) (Goodman et al., 2020)
35 which raises health concerns. Moreover, a range of products containing cannabis extracts
36 with low levels of THC, are sold legally and commercially (European Monitoring Centre
37 for Drugs and Drug Addiction (EMCDDA), 2021). Alongside these market changes, the
38 number of first-time cannabis treatment entrants is increasing (European Monitoring
39 Centre for Drugs and Drug Addiction (EMCDDA), 2021). Therefore, careful monitoring
40 of THC use is necessary to detect changes in consumption patterns and to understand the
41 shifts in the drug markets (Burgard et al., 2019). The most comprehensive approach
42 would thus consist of the triangulation of data from different sources e.g. key informants,
43 seizure data, population surveys, and city-based wastewater analysis.

44 Wastewater analysis of human biomarkers, also known as wastewater-based
45 epidemiology (WBE), has been an effective tool to show within- and between-week or
46 years patterns of drug use such as cocaine, MDMA, methamphetamine, and amphetamine
47 (Humphries et al., 2016; Ort et al., 2014). Moreover, the comparison of WBE data and
48 sales statistics has shown to be an accurate and complementary tool to estimate nicotine
49 and alcohol consumption (Lai et al., 2018). While the parent drugs were quantified for
50 methamphetamine, amphetamine, and MDMA, nicotine, ethanol, and cocaine

51 consumption were estimated by the quantification of their main human metabolites
52 (cotinine and hydroxy-cotinine; ethyl sulfate; and benzoylecgonine, respectively).
53 Regarding cannabis, THC is the major psychoactive ingredient, which is metabolized by
54 microsomal hydroxylation to the primary and intermediate metabolite, 11-hydroxy-THC
55 (THC-OH). Subsequently, THC-OH is further metabolized by the enzyme alcohol
56 dehydrogenase to 11-nor-9-carboxy-THC (THC-COOH), which is primarily quantified
57 in wastewater and used to estimate cannabis consumption (Bijlsma et al., 2020).

58 While WBE has been successfully implemented for the monitoring of stimulants
59 mentioned above, in the specific case of cannabis, the method suffers from various
60 inconsistencies (Burgard et al., 2019) and several studies have identified important
61 knowledge gaps related to the analytical determination of cannabis biomarkers in
62 wastewater (Causanilles et al., 2017a): (i) the possible sorption of the biomarkers to the
63 suspended solids in wastewater or to the biofilm of the sewer system (Ramin et al., 2017,
64 2016) and as a consequence the potential partition of the different biomarkers in the solid
65 and liquid phase of raw influent wastewater (IWW) (Ramin et al., 2017) and (ii) the
66 metabolism and excretion rates considering gender, race and routes of administration, and
67 subsequently the derived excretion factors (Khan and Nicell, 2012).

68 Results of inter-laboratory exercises accomplished by the Sewage Analysis CORE group
69 Europe (SCORE) revealed that, although laboratories were able to determine THC-
70 COOH in methanol successfully, its accurate determination in the liquid phase of IWW
71 was challenging (van Nuijs et al., 2018). Despite several improvements focusing on the
72 analytical procedure (Causanilles et al., 2017b), back-calculations of cannabis
73 consumption in WBE suggested important deviations from consumption estimates
74 obtained through conventional indicators (Bijlsma et al., 2021; Burgard et al., 2019;
75 Causanilles et al., 2017b). An important cause of these systematic deviations could be

76 related to the lower polarity of the cannabis biomarkers in comparison with other drugs
77 (e.g. pKa of THC ~ 10.6 and THC-COOH ~ 4.2), which would favour their sorption onto
78 suspended solids (Senta et al., 2013), as suggested by some authors (Burgard et al., 2019;
79 Khan and Nicell, 2012; Pandopulos et al., 2020a). Moreover, THC and its metabolites are
80 excreted via feces in a much higher proportion than other drugs (Gracia-Lor et al., 2016).
81 Hence, more emphasis needs to be placed on understanding cannabis biomarkers
82 distribution between the liquid and suspended solids fraction.

83 The aim of this work is to use different analytical approaches for the determination and
84 investigation of the three cannabis biomarkers (THC, THC-OH, and THC-COOH) in raw
85 IWW. The analytical determination of THC-COOH is commonly performed by liquid
86 chromatography tandem mass spectrometry (LC-MS/MS) with a previous sample
87 treatment consisting of a filtration or centrifugation step followed by pre-concentration
88 of the sample through solid-phase extraction (SPE) (Bijlsma et al., 2020, 2014;
89 Causanilles et al., 2017a). Other sample extraction alternatives are also reported, such as
90 liquid-liquid extraction (LLE) (González-Mariño et al., 2018; Pandopulos et al., 2020b;
91 Tschärke et al., 2016) and solid-phase microextraction (SPME) (Racamonde et al., 2012).
92 In one study using direct injection, THC-COOH was below the limit of detection in real
93 samples (Berset et al., 2010), suggesting that a concentration step was necessary. In this
94 research, both LLE and SPE were employed for the extraction of the liquid phase after
95 centrifugation, while the suspended solids were analyzed separately by solid-liquid
96 extraction (SLE). In parallel, the total raw IWW (without centrifugation or filtration) was
97 analyzed by LLE for comparison. This study allows a better understanding of the
98 occurrence (e.g. partition between the liquid phase and the suspended solids) of cannabis
99 biomarkers in IWW. Special attention was paid to THC-COOH as this is the biomarker
100 commonly used in WBE studies for estimating cannabis consumption.

101 2. Experimental

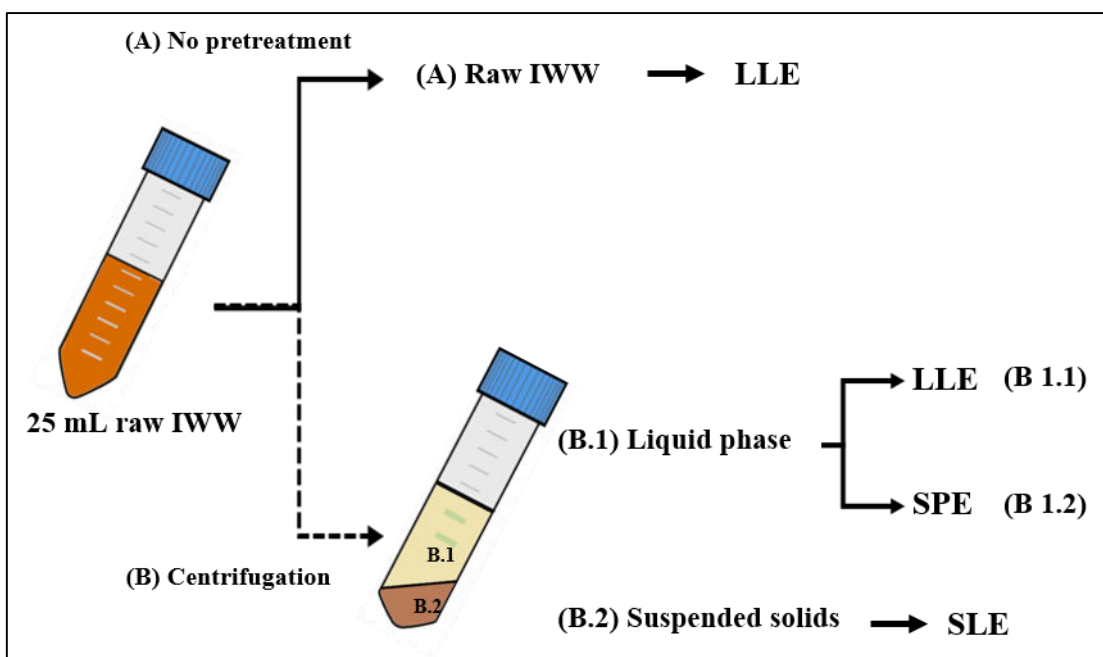
102 2.1. Chemicals and reagents

103 High purity analytical standards were purchased from Sigma-Aldrich (Cerilliant
104 Corporation, TX, USA). The standards used were Δ^9 -tetrahydrocannabinol (THC), 11-
105 hydroxy- Δ^9 -THC (THC-OH) and 11-nor-9-carboxy- Δ^9 -THC (THC-COOH) with their
106 respective isotope labelled internal standards (ILIS), THC-D₃, THC-OH-D₃ and THC-
107 COOH-D₃.

108 Individual standard stock solutions were prepared at 100 mg/L or 10 mg/L in methanol
109 (MeOH) and stored in amber glass vials at -20 °C. Multi-compound working solutions
110 were prepared by appropriate dilution of the standard stock solutions in MeOH. The
111 analytes working mix solution was prepared at 500 µg/L and the ILIS working mix
112 solution was prepared at 200 µg/L. LC-MS grade MeOH, hexane (HX), ethyl acetate
113 (EA), hydrochloric acid (HCl), formic acid (HCOOH), and sodium chloride (NaCl) were
114 supplied by Scharlab. HPLC-grade water was obtained by purifying demineralized water
115 using a Milli-Q system from Millipore (Bedford, MA, USA).

116 2.2. Sample collection and treatment

117 Influent wastewater samples (24-h composite, time-proportional with a time interval of
118 10 min) were collected from the wastewater treatment plant (WWTP) of Castellon, Spain.
119 After collection, the samples were immediately transported to the laboratory and stored
120 in the dark at -20 °C until analysis. Several extraction techniques were applied for the
121 sample treatment of entire raw IWW and the liquid phase, and the suspended solids
122 separately. These techniques included LLE, SPE, and SLE. **Figure 1** shows the different
123 extraction methods used.



124

125 **Figure 1.** Sample preparation for the analysis of the liquid and solid phase by different
 126 methods: liquid-liquid extraction (LLE), solid-phase extraction (SPE) and solid-liquid
 127 extraction (SLE).

128

129 2.2.1. Liquid-liquid extraction method

130 The LLE was applied for sample treatment of raw IWW (Figure 1, A) and the separated
 131 liquid phase (Figure 1, B 1.1). The sample was transferred to a 50 mL Falcon tube and 50
 132 μL of the ILIS working mix solution (200 $\mu\text{g}/\text{L}$) was added to 25 mL of non-centrifuged
 133 IWW (A) or centrifuged IWW (B.1), vortexed for 30 sec and let stand for two hours
 134 before extraction. Then, a spatula tip of NaCl was added and the sample was acidified to
 135 $\text{pH} \sim 2$ with HCl 1 M (400 μL), followed by 30 sec vortexing. Subsequently, 10 mL of
 136 HX:EA (2:1, v/v) was added, it was vortexed for 30 sec, sonicated 5 min and the content
 137 of the vessel was centrifuged at 5000 rpm for 5 min. A volume of 5 mL of the organic
 138 layer was transferred to a glass test tube and evaporated at 40 $^{\circ}\text{C}$ under a gentle stream of
 139 nitrogen. Extraction was executed once and the residue was reconstituted in a mixture of

140 300 μ L MeOH and 200 μ L Milli-Q water and the final extract was transferred to a vial
141 for LC-MS/MS analysis.

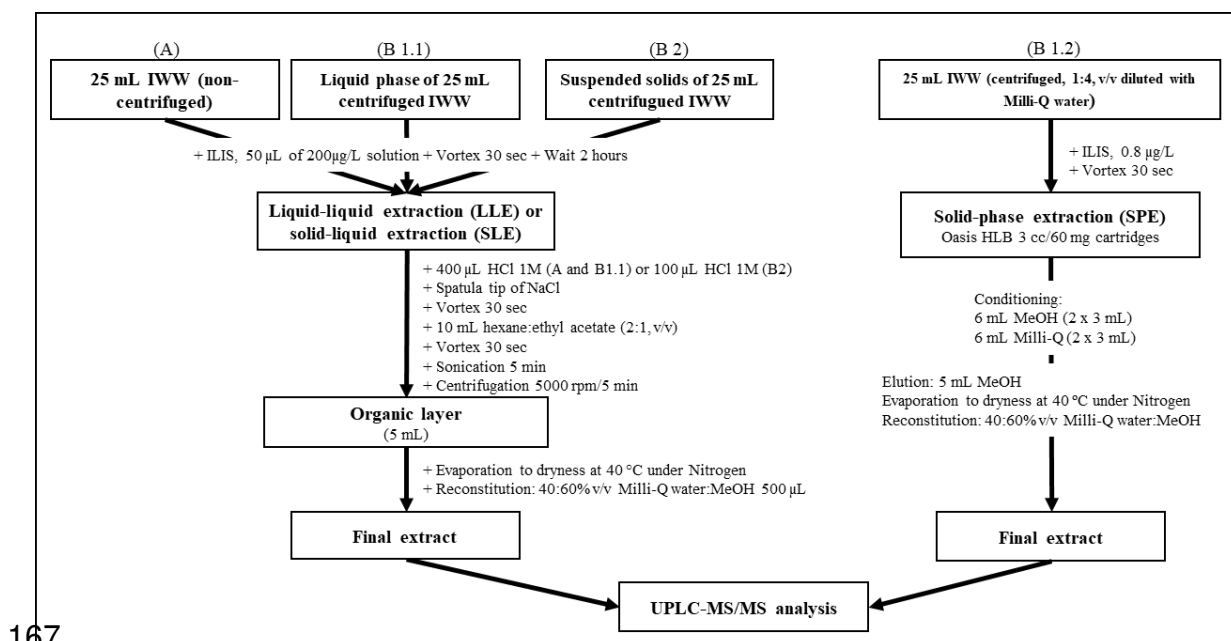
142 **2.2.2 Solid-liquid extraction method**

143 For the SLE of the analytes in the suspended solids (Figure 1, B.2), 25 mL of non-
144 centrifuged IWW was centrifuged at 5000 rpm for 5 min in a Falcon tube, the liquid phase
145 was removed and 50 μ L of the ILIS working mix solution (200 μ g/L) was added to the
146 suspended solids (pellet), vortexed for 30 sec and let stand for two hours before
147 extraction. Then, a spatula tip of NaCl and 100 μ L of HCl 1 M were added to the Falcon
148 tube and vortexed during 30 sec followed by adding 10 mL of HX:EA (2:1, v/v).
149 Subsequently, the content of the vessel was mixed by vortexing for 30 sec and sonicated
150 for 5 min. Finally, the content was centrifuged at 5000 rpm for 5 min, and 5 mL of the
151 organic layer was transferred to a glass test tube and evaporated at 40 °C under a gentle
152 stream of nitrogen. The residue was reconstituted in a mixture of 300 μ L MeOH and 200
153 μ L Milli-Q water, and the final extract was transferred to a vial for LC-MS/MS analysis.

154 **2.2.3 Solid-phase extraction method**

155 SPE applied to the separated liquid phase (Figure 1, B 1.2) was based on a previously
156 developed in-house method (Bijlsma et al., 2014). Briefly, 25 mL centrifuged IWW was
157 diluted with 75 mL Milli-Q water (leading to 100 mL of four-fold diluted centrifuged
158 IWW), and 50 μ L of the ILIS working mix solution (200 μ g/L) was added before SPE
159 with Oasis HLB cartridges (3 mL, 60 mg). The cartridges were conditioned by washing
160 and rinsing with 6 mL of MeOH and 6 mL of Milli-Q water. After conditioning, the
161 samples were percolated through the cartridges by gravity (flow rate of \sim 3 mL/min), and
162 vacuum dried for approximately 15 min. The analytes were eluted with 5 mL of MeOH
163 and the extract was evaporated to dryness at 40 °C under a gentle stream of nitrogen.
164 Finally, the residue was reconstituted in 0.5 mL Milli-Q water:MeOH: (40:60, v/v) and

165 transferred to a vial for LC-MS/MS analysis. A schematic overview of the sample
 166 treatment protocol is shown in **Figure 2**.



167

168 **Figure 2.** Graphical workflow of the analytical procedure liquid-liquid extraction
 169 (LLE), solid-liquid extraction (SLE) and solid-phase extraction (SPE).

170

171 2.3. Instrumentation

172 Ultra-high performance liquid chromatography coupled to tandem mass spectrometry
 173 (UHPLC-MS/MS) sample analysis was performed using a Waters Acquity H-Class
 174 UPLC system (Waters Corporation, MA, USA) coupled to a triple quadrupole mass
 175 spectrometer (Xevo TQS, Waters, Manchester, UK) equipped with an electrospray
 176 ionization source (ESI) operated in positive ionization mode. Chromatographic separation
 177 was carried out using an Acquity UPLC BEH C18 column (1.7 µm, 50 x 2.1 mm) from
 178 Waters at a flow rate of 0.3 mL/min. Column temperature was kept at 40 °C and the
 179 sample manager was kept at 10 °C. Mobile phase consisted of a gradient of A: Milli-Q
 180 water 0.01% HCOOH, 5 mM NH₄Ac and B: MeOH as follows: 0 min 60% B, 3.5 min

181 95% B, 5.0 min 95% B, 5.1 min 60% B until 7 min for re-equilibration of the column for
182 the next injection. Cone and desolvation gas flow were set to 250 L/h and 1200 L/h,
183 respectively. For the operation of MS/MS mode, collision gas was argon 99.995%
184 (Praxair, Madrid, Spain) set to 0.15 mL/min. The source temperature was kept at 150 °C,
185 desolvation temperature at 650 °C and capillary voltage was established at 1.5 kV. Dwell
186 times were established at 15 ms. Selected transitions, cone voltages and collision energies
187 can be observed in **Table S-1**. All data were acquired and processed using MassLynx
188 v4.1 software (Waters, Manchester, UK).

189 **2.5. Method validation**

190 Method performance was evaluated with authentic IWW samples in terms of linearity,
191 limits of detection (LODs), limits of quantification (LOQs), accuracy (in terms of
192 recovery), and precision (inter-day precision expressed as relative standard deviation
193 (RSD)) taking into account the SANTE guideline (SANTE/12682/2019, 2019). Linearity
194 was studied by the preparation of calibration curves, using linear regression ($r^2 > 0.9900$)
195 with concentrations ranging from 50 to 100000 ng/L. LODs and LOQs were estimated
196 by analyzing spiked IWW at 100 ng/L based on a signal-to-noise (S/N) ratio of 3 and 10,
197 respectively. Accuracy and intra-day precision were evaluated using spiked IWW
198 samples ($n=5$, from different origin) at two concentration levels (100 and 800 ng/L)
199 quantified after ILIS correction. Recoveries were considered satisfactory when they
200 ranged between 60% and 120%, with RSD values lower than 20% (SANTE/12682/2019,
201 2019). Due to the impossibility of obtaining real “blank IWW samples”, as all of the target
202 analytes are usually present in IWW, samples were initially analyzed without the spiking
203 of the analytes and the quantified amount of the analytes was subtracted from the
204 measured concentration in spiked IWW.

205

206 **2.6. Stability experiments**

207 The in-sample stability of THC, THC-OH and THC-COOH was tested at three
208 temperatures (-20 °C, 4 °C and 20 °C) over 30 days at 0 h, 6 h, 12 h, 24 h, 3 d, 7 d, 14 d,
209 and 30 d. For each storage temperature, 2 bottles of 100 mL of non-centrifuged IWW
210 (one “blank” and one spiked at 1 µg/L with a mix of the three analytes) and 2 bottles of
211 100 mL of centrifuged IWW (one “blank” and one spiked at 1 µg/L with a mix of the
212 three analytes) were prepared and the ILIS mix solution was added in all bottles at 1 µg/L.
213 Then, samples were homogenized and distributed in 96 conical tubes (12 for each time
214 frame), containing 10 mL of sample. LLE was performed for the non-centrifuged and
215 centrifuged IWW at the three temperatures tested. For experiments at -20 °C, thawing was
216 done by adding mechanical shaking and heating with human temperature (holding in
217 hands). After LLE, the extracts were stored at -20 °C in a vial until LC-MS/MS analysis.
218 **Figure S-1** shows the procedure applied in the stability experiments.

219

220

221 3. Results and discussion

222 3.1. Stability experiments

223 The results obtained in the in-sample stability experiments are summarized in **Figures**
224 **S2-S4**. Concentration at time zero is considered as 100% of recovery. In the case of non-
225 centrifuged IWW, the three compounds were generally stable at -20 °C, 4 °C, and 20 °C
226 for up to one month as reported by *Causanilles et al* (Causanilles et al., 2017b).
227 Oppositely, a slight increase in recovery (up to 140%) was observed for THC-COOH at
228 all temperatures (**Figure S-2**). Desorption of THC-COOH present in the suspended solids
229 is the most probable reason for the increase in concentration, yet an interconversion of
230 compounds, due to the transformation of THC-OH to THC-COOH by oxidation might
231 also occur although less likely (Ramin et al., 2017). However, the obtained data cannot
232 support any of these hypotheses since a mixed spiking solution was used, and the real
233 solid used is not an authentic blank due to the presence of all three compounds on it.

234 The stability data of the analytes in centrifuged IWW presented more variability. All
235 compounds were stable up to 1 month at 4 °C and -20 °C, as reported previously
236 (Causanilles et al., 2017a; González-Mariño et al., 2018; Heuett et al., 2015). However,
237 in the case of THC-COOH (**Figure S-2**) and THC-OH (**Figure S-3**), notable losses were
238 observed when stored at 20 °C after two weeks. These results illustrate the relevance of
239 appropriate storage conditions of the samples, with the recommendation of storing the
240 samples at -20 °C, if analysis cannot be performed within 14 days after sample reception.
241 If analysis is performed within 14 days, the samples could be stored at 4 °C without
242 significant loss of analytes. Anyway, further research is necessary by spiking samples
243 individually with each analyte to clarify the possible interconversion mentioned above.

244

245 3.2. Liquid-liquid extraction of raw IWW

246 Previous publications reporting the use of LLE for THC-COOH extraction from IWW
247 (González-Mariño et al., 2018; Pandopulos et al., 2020b; Tschärke et al., 2016) were used
248 as a guide in the performed study. The tested solvent mixture was HX and EA, since HX
249 has been reported as appropriate to extract THC, and EA or HX:EA (1:1, v/v) to extract
250 THC-OH and THC-COOH from ultra-pure water (González-Mariño et al., 2018). EA has
251 also been reported for extraction of THC-COOH from wastewater (Pandopulos et al.,
252 2020b). In the present work, HX:EA (2:1, v/v) was chosen as the extraction solvent, as it
253 resulted in the best recoveries (**Table S-2**).

254 The addition of NaCl to the sample was also evaluated. Although no significant
255 differences were found in the recovery of analytes from IWW, the addition of NaCl was
256 eventually applied since IWW samples are highly variable, and previous research
257 recommended the addition of NaCl to improve recovery by “salting-out” target analytes
258 and to prevent the formation of emulsion in the LLE process (Causanilles et al., 2017b;
259 Pandopulos et al., 2020a). The waiting time between the addition of ILIS to the IWW and
260 the addition of HCl (1 M) was also evaluated at 20 min, 2 h, and overnight (14 h),
261 obtaining the most reproducible results when the ILIS was added and let stand minimum
262 for 2 h before starting the LLE process.

263 Data obtained in the validation of the LLE procedure applied to raw IWW (non-
264 centrifuged) are shown in **Table 1**. It was not possible to obtain real “blank” IWW,
265 because of the frequent occurrence of these three biomarkers in wastewater. This fact
266 impacted the validation process, especially at low analyte concentrations (*i.e.*, at 100 ng/L
267 spiked level, which was similar or even lower than the concentration of the analyte present
268 in the “blank” IWW used for validation). Recoveries were around 70% for the three
269 compounds with low RSDs ($\leq 10\%$), and all analytes could be fully identified in the

270 sample with two confirmatory transitions (q1, q2) and low deviations ($\leq 28\%$) in the q/Q
 271 ratios in relation to the reference standard average values.

272

273 **Table 1.** Liquid-liquid extraction method validation in raw influent wastewater (n=5).

Compound	LOQ (ng/L)	LOD (ng/L)	Conc in "blank" (ng/L)	Recovery, %		q1/Q ratio deviation (%)		q2/Q ratio deviation (%)			
				(RSD)		100	800	100	800	100	800
				ng/L	ng/L	ng/L	ng/L	ng/L	ng/L		
THC	10	3	51	*	65 (4)	3	7	28	1		
THC-OH	5	2	104	*	80 (9)	3	0.3	4	3		
THC-COOH	3	1	246	*	73 (10)	10	7	1	2		

274 *Not estimated due to the high concentration of the analyte in the spiked "blank" sample.

275

276 3.3. Analysis of the liquid phase

277 In this study the liquid phase of the centrifuged IWW samples was extracted using both
 278 LLE and SPE separately and the performance of the two approaches was compared
 279 (**Figure 2**). In the case of the SPE, two sorbents *i.e.*, Strata X (60 mg, 3 mL) and Oasis
 280 HLB (60 mg, 3 mL) often applied in multi residue methods, were tested for non-spiked
 281 and spiked IWW samples. Oasis HLB cartridges led to good recoveries (82 - 130%) and
 282 were selected for subsequent experiments (**Table S-3**). In parallel, LLE was also tested
 283 for the extraction of the liquid phase of IWW, and both procedures were finally validated
 284 (**Table 2**).

285

286 **Table 2.** Liquid-liquid extraction and solid-phase extraction method validation in the
 287 liquid phase of influent wastewater (n=3).

Compound	LOQ (ng/L)	LOD (ng/L)	Concentration in “blank” (ng/L) ^(a)	Recovery, % (RSD)		q1/Q ratio deviation (%)		q2/Q ratio deviation (%)	
				100	800	100	800	100	800
				ng/L	ng/L	ng/L	ng/L	ng/L	ng/L
THC	10	3	-	70 (6)	72 (3)	7	6	9	4
LLE THC-OH	5	2	35	71 (4)	78 (5)	5	3	2	1
THC-COOH	3	1	183	*	69 (6)	10	8	3	1
THC	20	6	-	84 (9)	82 (4)	5	9	14	22
SPE THC-OH	12	4	35	95 (3)	90 (1)	22	19	7	10
THC-COOH	26	8	183	*	98 (6)	4	5	26	13

288 *Not estimated due to the high concentration of the analyte in the “blank” sample.

289 ^(a)Average value of the “blank” concentration obtained by SPE and LLE extraction methods.

290

291 Accuracy was consistently below 100%, with SPE recoveries being slightly higher at both
 292 validated levels. Precision was satisfactory, with RSD ≤ 10% in all cases. LOQs (from 3
 293 to 10 ng/L in the case of LLE and from 12 to 26 ng/L in the SPE) and LODs were lower
 294 in LLE for all compounds studied (**Table 2**).

295 In order to obtain more data to compare both procedures, seven consecutive samples (*i.e.*,
 296 in whole week) were processed using both methods. Concentrations of THC-COOH
 297 showed deviations < 30% in 6 out of the 7 samples analyzed when comparing data from
 298 both methods (**Table 3**). Despite the, in general, slightly higher recoveries when
 299 employing SPE, the LLE procedure (25 mL of sample extracted with 10 mL of HX:EA
 300 (2:1, v/v)) was considered as a good alternative for the analysis of these compounds,
 301 taking into account the higher cost of SPE and the more time-consuming steps (*i.e.*
 302 conditioning, sample loading, washing, and elution).

303

304 **Table 3.** Determination of THC-COOH by liquid-liquid extraction and solid-phase
305 extraction in seven centrifuged influent wastewater samples from a one-week sampling.

Sample	THC-COOH (ng/L)		Deviation SPE/LLE (%)
	LLE	SPE	
IWW 1	239	336	+41
IWW 2	270	328	+22
IWW 3	262	308	+18
IWW 4	221	208	-6
IWW 5	327	348	+6
IWW 6	281	356	+27
IWW 7	332	308	-7

306

307 **3.4. Analysis of suspended solids**

308 The SLE method for suspended solids was tested in terms of extraction system, solvent,
309 and extraction time. Three different extraction systems were tested, including vortex-
310 assisted (1 min), rotatory-assisted (time=10, 20 and 30 min), and ultrasonic-assisted
311 extraction (time=10, 20 and 30 min). Based on the data summarized in **Table S-4**,
312 ultrasonic-assisted extraction for 10 min led to the highest extraction of cannabis
313 biomarkers and therefore it was chosen for subsequent experiments. Next, ultrasonic-
314 assisted extraction was performed with different ratios of HX:EA, including 1:1 (v/v), 2:1
315 (v/v), 3:1 (v/v), and 1:2 (v/v), all tested at different time frames (t = 2, 5 and 10 min). The
316 best results in terms of extraction efficiency were obtained with HX:EA (2:1, v/v) during
317 5 min, which was finally selected as the optimal procedure for the extraction of cannabis
318 biomarkers from suspended solids (**Table S-5**).

319 The validation of the SLE procedure was subjected to practical challenges, because of
320 difficulties to accurately weigh or measure the amount of solid in each sample aliquot
321 used for validation. To try to overcome this situation, the samples subjected to validation
322 were shaken vigorously until all solid particles were floating homogeneously in the liquid

323 and subsequently 25 mL aliquots were collected and centrifuged. After that, the liquid
 324 phase was removed as much as possible, leaving the pellet (suspended solids) at the
 325 bottom of the Falcon tube to proceed with the extraction and validation. In this way, RSDs
 326 below 20% were obtained in all cases, and recoveries were close to 100% for the three
 327 cannabis biomarkers (**Table 4**).

328

329 **Table 4.** Solid-liquid extraction method validation in the suspended solids of 25 mL
 330 influent wastewater (n=3).

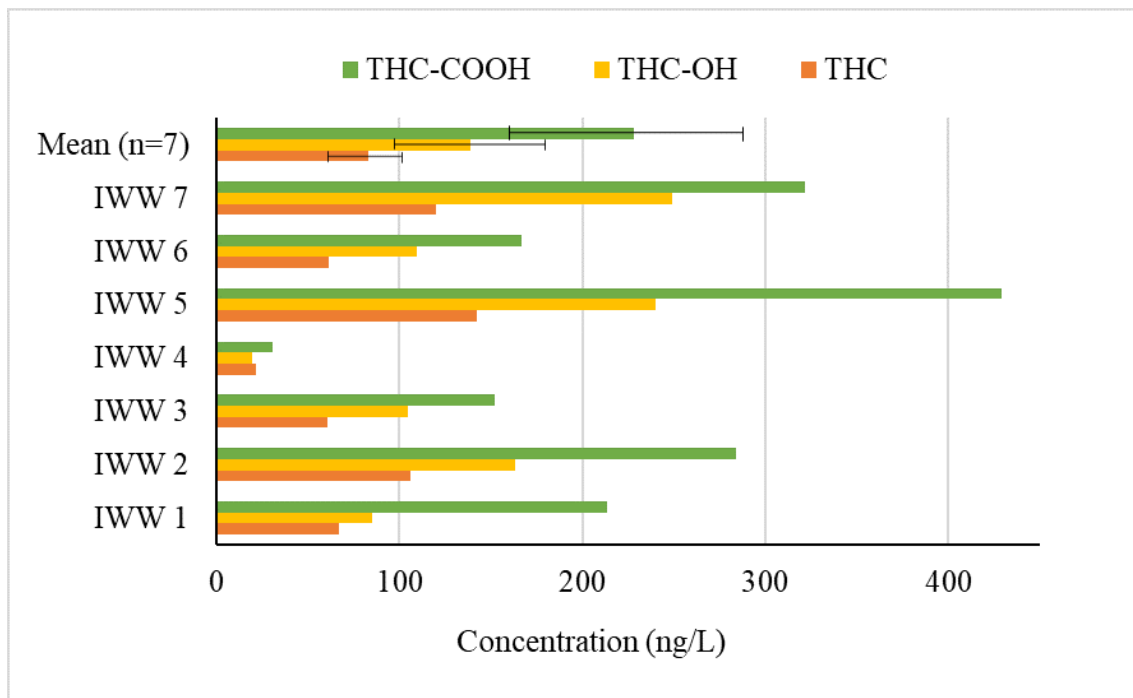
Compound	LOQ (ng)	LOD (ng)	Concentration in "blank" (ng/L) ^(a)	Recoveries at 20 ng, % (RSD)
THC	0.22	0.07	79	97 (12)
THC-OH	0.21	0.07	175	101 (16)
THC-COOH	0.18	0.06	338	106 (15)

331 ^(a) Calculated from the mass extracted from the entire pellet and the volume of raw
 332 IWW sample (25 mL).

333

334 This method was applied to the suspended solids of the seven samples mentioned in
 335 Section 3.3 of which the results are shown in **Figure 3**. The amount of biomarker
 336 quantified in the suspended solids of 25 mL of IWW was converted in ng/L. It can be
 337 seen that the three cannabis metabolites were present in all samples, the predominant
 338 compound being THC-COOH, followed by THC-OH and THC. These findings are in
 339 agreement with the low polarity of the compounds, which are consequently substantially
 340 sorbed onto the solid phase of IWW. It should be noted that low concentrations,
 341 particularly for THC-OH and THC-COOH, were found in the solids of sample IWW4,
 342 which was characterized by a low content of suspended solids (visual observation). These

343 two values were confirmed as outliers (test Q Dixon) and were removed to obtain the
344 average of each biomarker present in the suspended solids.



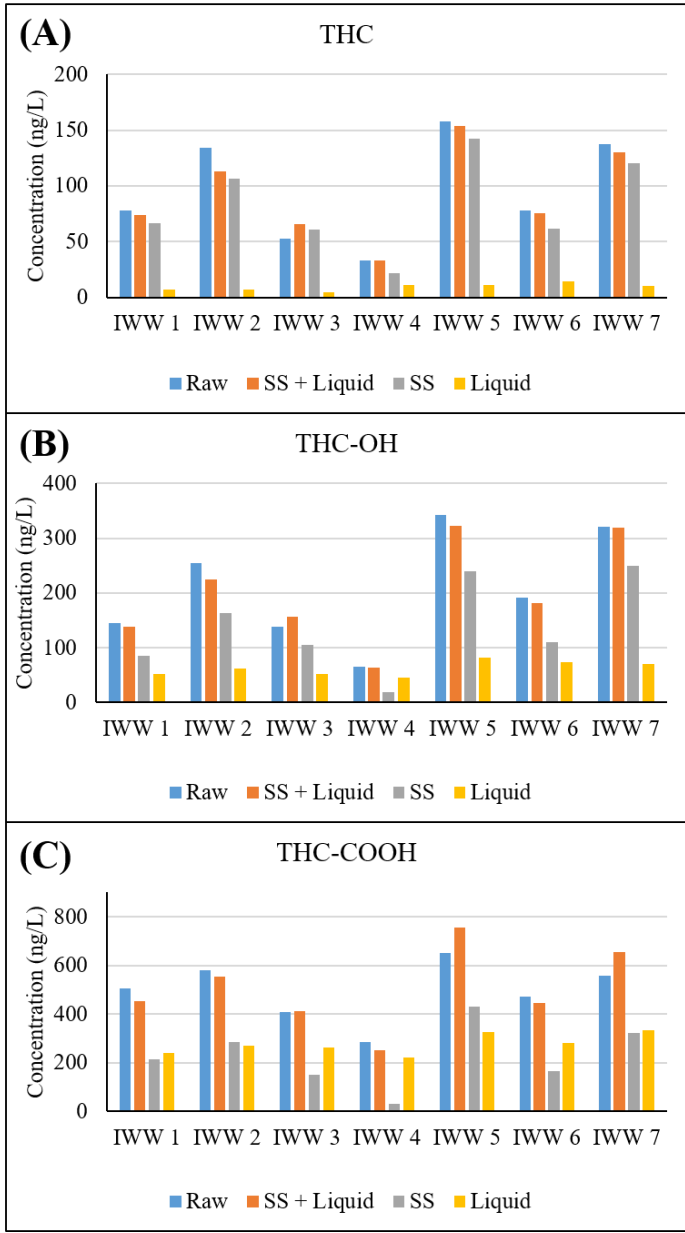
345
346 **Figure 3.** Cannabis biomarkers in the suspended solids of seven influent wastewater
347 samples.

348 It can be concluded that analyzing only the liquid phase of IWW (e.g. after centrifugation
349 or filtration), independently of whether SPE or LLE is applied for that analysis, would
350 imply that only a fraction of the cannabis biomarkers is measured. The obtained results
351 indicate that an important amount of cannabis biomarkers is present in the suspended
352 solids, a fact that should be taken into account when performing wastewater analysis on
353 cannabis biomarkers.

354 3.5. Analysis of raw IWW

355 The seven IWW samples under investigation were also analyzed by LLE without previous
356 centrifugation (i.e. analysis of the raw IWW including the liquid phase and suspended
357 solids) (**Figure 4**). This allows to compare the total amount of biomarkers obtained by
358 LLE of the whole raw wastewater with the sum of the suspended solids and liquid phase

359 biomarkers analyzed separately (data given in previous sections). The obtained data show
 360 a good agreement using both approaches (deviation <5% in 70% of the results and <20%
 361 in the remaining data). These results support the hypothesis that data obtained analyzing
 362 the raw IWW by LLE without previous removal of the suspended solids, are rather similar
 363 to the sum of biomarkers in the suspended solids and in the liquid phase (either extracted
 364 by SPE or LLE) analyzed separately.



365
 366 **Figure 4.** Analysis of (A) THC, (B) THC-OH and (C) THC-COOH in different phases
 367 of influent wastewater. Comparison of the whole raw influent wastewater (blue bar), the

368 liquid phase (yellow bar), the suspended solids (SS) (grey bar) and the sum of SS and the
369 liquid phase analyzed separately (orange bar).

370 Taking into consideration the data obtained in this study, two approaches could be
371 implemented in future studies to further improve the knowledge on cannabis biomarker
372 concentrations in IWW:

373 (i) to analyze the raw IWW by LLE without separating the sample into the liquid and
374 solid phase by filtration or centrifugation. This would imply the use of an extra aliquot of
375 the sample and a dedicated LLE extraction procedure, in addition to that used for
376 conventional analysis of other illicit drugs that it is normally based on SPE. The limited
377 information available in relation to urinary and fecal excretion rates and representative
378 sampling for solids are currently bottlenecks and, therefore this issue should be studied
379 in more depth. In this context, a parameter such as turbidity to evaluate the suspended
380 solids in the sample collected could give more insight.

381 (ii) to perform analysis of the liquid phase of IWW by either SPE or LLE (following
382 centrifugation/filtration of the sample to remove the suspended solids). As this procedure
383 does not include the fraction in the solid phase, a correction factor could be applied to the
384 measured concentration in order to provide a better estimation of the total biomarker
385 concentration in raw IWW. However, the suitability and robustness of such a correction
386 factor should be evaluated. This would include the analysis of a large number of
387 wastewater samples collected from different locations with different composition and
388 characteristics to assess spatial and temporal variation. These experiments will be
389 performed in the near future as this approach appears as a good option for most multi-
390 residue, multi-class analysis, where THC-COOH is determined together with other illicit
391 drugs following a common sample treatment (typically SPE);

392 Despite the important contribution of suspended solids to the total measured analytes in
393 the sample, there is still a knowledge gap in the possible adsorption or desorption
394 processes that may exist between the suspended solids and the liquid phase. Preliminary
395 sorption experiments have been performed in Milli-Q water containing an amount of
396 suspended solids, but no conclusive results have been obtained until now. Further
397 research is planned applying the use of THC-COOH-D₉ as an analogue to study the
398 behaviour and distribution of THC-COOH between both phases in different IWW.

399

400 **4. Conclusions**

401 In this research, an analytical method has been developed for cannabis biomarkers in
402 wastewater, with a focus on their occurrence and distribution in the liquid phase and
403 suspended solids. Data from this paper show that LLE of raw IWW allowed obtaining
404 cannabis biomarker concentrations in both the liquid and solid phase. The separate
405 analysis of the liquid and solid phase (e.g. after centrifugation of raw IWW) revealed that
406 a high percentage of the compounds present in influent wastewater corresponded to the
407 solid phase (average 90% THC, 69% THC-OH, 42% THC-COOH). To date, the most
408 common analytical protocol for cannabis biomarkers analysis in IWW, consisting of the
409 application of SPE, only considers the liquid phase and do not consider their presence in
410 the suspended solids. This, consequently, leads to an underestimation of the total
411 biomarker amount present in IWW. Thus, the analysis of the IWW without separation of
412 the solid phase by LLE offers more realistic information on the biomarker concentration
413 in IWW than analysis of only the liquid phase. Hence, a better comprehension in the
414 complexity of measuring cannabis biomarkers in IWW is given. Moreover, the
415 preliminary results allowed to identify the need for future research where the following
416 points should be addressed i) assess sampling uncertainty related to solids ii) partition of
417 the cannabis biomarkers between liquid and solid phases during in-sewer transport and
418 sample storage, and iii) obtaining accurate urinary and fecal excretion rates. By answering
419 these knowledge gaps more insight will be obtained in how to use WBE as a tool to
420 monitor cannabis consumption.

421

422 **Acknowledgments**

423 The authors acknowledge the financial support by the European Union's Justice
424 Programme — Drugs Policy Initiatives EuSeME (project number 861602). Ph.D. M.C.

425 Campos-Mañas acknowledges MINECO for her Juan de la Cierva-formación grant (Ref.
426 FJC-2019-039912-I). IUPA team from Univ Jaume I acknowledges the financial support
427 of Generalitat Valenciana as a research group of excellence (Prometeo 2019/040).

428 **References**

- 429 Berset, J.D., Brenneisen, R., Mathieu, C., 2010. Analysis of illicit and illicit drugs in
430 waste, surface and lake water samples using large volume direct injection high
431 performance liquid chromatography - Electrospray tandem mass spectrometry
432 (HPLC-MS/MS). *Chemosphere* 81, 859–866.
433 <https://doi.org/10.1016/j.chemosphere.2010.08.011>
- 434 Bijlsma, L., Beltrán, E., Boix, C., Sancho, J. V., Hernández, F., 2014. Improvements in
435 analytical methodology for the determination of frequently consumed illicit drugs in
436 urban wastewater. *Anal. Bioanal. Chem.* 406, 4261–4272.
437 <https://doi.org/10.1007/s00216-014-7818-4>
- 438 Bijlsma, L., Burgard, D.A., Been, F., Ort, C., Matias, J., Yargeau, V., 2020. The
439 estimation of cannabis consumption through wastewater analysis, in:
440 *Comprehensive Analytical Chemistry*. Elsevier B.V., pp. 453–482.
441 <https://doi.org/10.1016/bs.coac.2020.04.005>
- 442 Bijlsma, L., Picó, Y., Andreu, V., Celma, A., Estévez-Danta, A., González-Mariño, I.,
443 Hernández, F., López de Alda, M., López-García, E., Marcé, R.M., Miró, M.,
444 Montes, R., Pérez de San Román-Landa, U., Pitarch, E., Pocurull, E., Postigo, C.,
445 Prieto, A., Rico, A., Rodil, R., Valcárcel, Y., Ventura, M., Quintana, J.B., 2021. The
446 embodiment of wastewater data for the estimation of illicit drug consumption in
447 Spain. *Sci. Total Environ.* 772. <https://doi.org/10.1016/j.scitotenv.2020.144794>
- 448 Burgard, D.A., Williams, J., Westerman, D., Rushing, R., Carpenter, R., LaRock, A.,
449 Sadetsky, J., Clarke, J., Fryhle, H., Pellman, M., Banta-Green, C.J., 2019. Using
450 wastewater-based analysis to monitor the effects of legalized retail sales on cannabis
451 consumption in Washington State, USA. *Addiction* 114, 1582–1590.

452 <https://doi.org/10.1111/add.14641>

453 Causanilles, A., Baz-Lomba, J.A., Burgard, D.A., Emke, E., González-Mariño, I.,
454 Krizman-Matasic, I., Li, A., Löve, A.S.C., McCall, A.K., Montes, R., van Nuijs,
455 A.L.N., Ort, C., Quintana, J.B., Senta, I., Terzic, S., Hernandez, F., de Voogt, P.,
456 Bijlsma, L., 2017a. Improving wastewater-based epidemiology to estimate cannabis
457 use: focus on the initial aspects of the analytical procedure. *Anal. Chim. Acta* 988,
458 27–33. <https://doi.org/10.1016/j.aca.2017.08.011>

459 Causanilles, A., Baz-Lomba, J.A., Burgard, D.A., Emke, E., González-Mariño, I.,
460 Krizman-Matasic, I., Li, A., Löve, A.S.C., McCall, A.K., Montes, R., van Nuijs,
461 A.L.N., Ort, C., Quintana, J.B., Senta, I., Terzic, S., Hernandez, F., de Voogt, P.,
462 Bijlsma, L., 2017b. Improving wastewater-based epidemiology to estimate cannabis
463 use: focus on the initial aspects of the analytical procedure. *Anal. Chim. Acta* 988,
464 27–33. <https://doi.org/10.1016/j.aca.2017.08.011>

465 European Monitoring Centre for Drugs and Drug Addiction (EMCDDA), 2021. European
466 Drug Report 2021: Trends and Developments.

467 European Monitoring Centre for Drugs and Drug Addiction (EMCDDA), 2017. Cannabis
468 legislation in Europe, Publications Office of the European Union, Luxembourg.
469 <https://doi.org/10.2810/930744>

470 González-Mariño, I., Thomas, K. V., Reid, M.J., 2018. Determination of cannabinoid and
471 synthetic cannabinoid metabolites in wastewater by liquid–liquid extraction and
472 ultra-high performance supercritical fluid chromatography-tandem mass
473 spectrometry. *Drug Test. Anal.* 10, 222–228. <https://doi.org/10.1002/dta.2199>

474 Goodman, S., Wadsworth, E., Leos-Toro, C., Hammond, D., 2020. Prevalence and forms
475 of cannabis use in legal vs. illegal recreational cannabis markets. *Int. J. Drug Policy*

476 76, 102658. <https://doi.org/10.1016/j.drugpo.2019.102658>

477 Gracia-Lor, E., Zuccato, E., Castiglioni, S., 2016. Refining correction factors for back-
478 calculation of illicit drug use. *Sci. Total Environ.* 573, 1648–1659.
479 <https://doi.org/10.1016/j.scitotenv.2016.09.179>

480 Heuett, N. V., Ramirez, C.E., Fernandez, A., Gardinali, P.R., 2015. Analysis of drugs of
481 abuse by online SPE-LC high resolution mass spectrometry: Communal assessment
482 of consumption. *Sci. Total Environ.* 511, 319–330.
483 <https://doi.org/10.1016/j.scitotenv.2014.12.043>

484 Humphries, M.A., Bruno, R., Lai, F.Y., Thai, P.K., Holland, B.R., O'Brien, J.W., Ort, C.,
485 Mueller, J.F., 2016. Evaluation of Monitoring Schemes for Wastewater-Based
486 Epidemiology to Identify Drug Use Trends Using Cocaine, Methamphetamine,
487 MDMA and Methadone. *Environ. Sci. Technol.* 50, 4760–4768.
488 <https://doi.org/10.1021/acs.est.5b06126>

489 Khan, U., Nicell, J.A., 2012. Sewer epidemiology mass balances for assessing the illicit
490 use of methamphetamine, amphetamine and tetrahydrocannabinol. *Sci. Total*
491 *Environ.* 421–422, 144–162. <https://doi.org/10.1016/j.scitotenv.2012.01.020>

492 Lai, F.Y., Gartner, C., Hall, W., Carter, S., O'Brien, J., Tschärke, B.J., Been, F., Gerber,
493 C., White, J., Thai, P., Bruno, R., Prichard, J., Kirkbride, K.P., Mueller, J.F., 2018.
494 Measuring spatial and temporal trends of nicotine and alcohol consumption in
495 Australia using wastewater-based epidemiology. *Addiction* 113, 1127–1136.
496 <https://doi.org/10.1111/add.14157>

497 Ort, C., Eppler, J.M., Scheidegger, A., Rieckermann, J., Kinzig, M., Sörgel, F., 2014.
498 Challenges of surveying wastewater drug loads of small populations and
499 generalizable aspects on optimizing monitoring design. *Addiction* 109, 472–481.

500 <https://doi.org/10.1111/add.12405>

501 Pandopulos, A.J., Bade, R., O'Brien, J.W., Tschärke, B.J., Mueller, J.F., Thomas, K.,
502 White, J.M., Gerber, C., 2020a. Towards an efficient method for the extraction and
503 analysis of cannabinoids in wastewater. *Talanta* 217.
504 <https://doi.org/10.1016/j.talanta.2020.121034>

505 Pandopulos, A.J., Bade, R., O'Brien, J.W., Tschärke, B.J., Mueller, J.F., Thomas, K.,
506 White, J.M., Gerber, C., 2020b. Towards an efficient method for the extraction and
507 analysis of cannabinoids in wastewater. *Talanta* 217, 121034.
508 <https://doi.org/10.1016/j.talanta.2020.121034>

509 Racamonde, I., Villaverde-de-Sáa, E., Rodil, R., Quintana, J.B., Cela, R., 2012.
510 Determination of Δ^9 -tetrahydrocannabinol and 11-nor-9-carboxy- Δ^9 -
511 tetrahydrocannabinol in water samples by solid-phase microextraction with on-fiber
512 derivatization and gas chromatography-mass spectrometry. *J. Chromatogr. A* 1245,
513 167–174. <https://doi.org/10.1016/j.chroma.2012.05.017>

514 Ramin, P., Brock, A.L., Causanilles, A., Valverde-Pérez, B., Emke, E., De Voogt, P.,
515 Polesel, F., Plósz, B.G., 2017. Transformation and Sorption of Illicit Drug
516 Biomarkers in Sewer Biofilms. *Environ. Sci. Technol.* 51, 10572–10584.
517 <https://doi.org/10.1021/acs.est.6b06277>

518 Ramin, P., Brock, A.L., Polesel, F., Causanilles, A., Emke, E., De Voogt, P., Plosz, B.G.,
519 2016. Transformation and sorption of illicit drug biomarkers in sewer systems:
520 Understanding the role of suspended solids in raw wastewater. *Environ. Sci.*
521 *Technol.* 50, 13397–13408. <https://doi.org/10.1021/acs.est.6b03049>

522 SANTE/12682/2019, 2019. Guidance document on analytical quality control and method
523 validation for pesticide residues analysis in food and feed. *Saf. Food Chain Pestic.*

524 Biocides. Eur. Comm. 1–48.

525 Senta, I., Krizman, I., Ahel, M., Terzic, S., 2013. Integrated procedure for multiresidue
526 analysis of dissolved and particulate drugs in municipal wastewater by liquid
527 chromatography-tandem mass spectrometry. Anal. Bioanal. Chem. 405, 3255–3268.
528 <https://doi.org/10.1007/s00216-013-6720-9>

529 Tscharke, B.J., Chen, C., Gerber, J.P., White, J.M., 2016. Temporal trends in drug use in
530 Adelaide, South Australia by wastewater analysis. Sci. Total Environ. 565, 384–391.
531 <https://doi.org/10.1016/j.scitotenv.2016.04.183>

532 United Nations Office on Drugs and Crime [WWW Document], n.d. URL
533 <https://www.unodc.org/unodc/en/data-and-analysis/wdr2021.html>

534 van Nuijs, A.L.N., Lai, F.Y., Been, F., Andres-Costa, M.J., Barron, L., Baz-Lomba, J.A.,
535 Berset, J.D., Benaglia, L., Bijlsma, L., Burgard, D., Castiglioni, S., Christophoridis,
536 C., Covaci, A., de Voogt, P., Emke, E., Fatta-Kassinos, D., Fick, J., Hernandez, F.,
537 Gerber, C., González-Mariño, I., Grabic, R., Gunnar, T., Kannan, K., Karolak, S.,
538 Kasprzyk-Hordern, B., Kokot, Z., Krizman-Matasic, I., Li, A., Li, X., Löve, A.S.C.,
539 Lopez de Alda, M., McCall, A.K., Meyer, M.R., Oberacher, H., O'Brien, J.,
540 Quintana, J.B., Reid, M., Schneider, S., Simoes, S.S., Thomaidis, N.S., Thomas, K.,
541 Yargeau, V., Ort, C., 2018. Multi-year inter-laboratory exercises for the analysis of
542 illicit drugs and metabolites in wastewater: Development of a quality control system.
543 TrAC - Trends Anal. Chem. 103, 34–43. <https://doi.org/10.1016/j.trac.2018.03.009>

544



Click here to access/download

**Electronic Supplementary Material (for online publication
only)**

Supplementary Data.pdf

